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Specific and generic stem biomass and volume models of tree species in a West African tropical semi-deciduous forest

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Highlights

- Non-destructive sampling approach applied to derive ground truth observations and generate robust basic wood densities.
- Species-specific and generic allometric equations.
- Specific equations have better predictive capabilities than generic models.

Abstract

The quantification of the contribution of tropical forests to global carbon stocks and climate change mitigation requires availability of data and tools such as allometric equations. This study made available volume and biomass models for eighteen tree species in a semi-deciduous tropical forest in West Africa. Generic models were also developed for the forest ecosystem, and basic wood density determined for the tree species. Non-destructive sampling approach was carried out on five hundred and one sample trees to analyse stem volume and biomass. From the modelling of volume and biomass as functions of diameter at breast height (Dbh) and stem height, logarithmic models had better predictive capabilities. The model validation showed that in absence of data on height, models using Dbh only as variable was an alternative. The comparison of basic wood densities to data published in literature enabled to conclude that the non-destructive sampling was a good approach to determining reliable basic wood density. The comparative analysis of species-specific models in this study with selected generic models for tropical forests indicated low probability to identify effective generic models with good predictive ability for biomass. Given tree species richness of tropical forests, the study demonstrated the hypothesis that species-specific models are preferred to generic models, and concluded that further research should be oriented towards development of specific models to cover the full range of dominant tree species of African forests.

Keywords non-destructive sampling; basic wood density; allometric equations; carbon stock

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1 Introduction

Human activities such as commercial fellings, fuelwood gathering, fires, and agricultural activities disturb forest ecosystems through deforestation and forest degradation and release the carbon stored in living biomass, soil, dead organic matter and litter pools to the atmosphere. Between the years 2000 and 2010, more than 3 million ha of tropical forest were lost in Africa including 875 000 ha in West Africa (FAO 2011). To address the issue a number of human-induced activities including reducing emissions from deforestation and forest degradation, sustainable management of forests, conservation and enhancement of forest carbon stocks have been proposed by the international community (UNFCCC 2010). To assess the performance of these activities, in terms of emissions reduction compared to a reference level, historical and current data on carbon stocks as well as tools such as biomass models would be required.

Databases on biomass and volume functions have been developed in other regions including Europe (Zianis et al. 2005), North America (Tritton and Hornbeck 1982; Ter-Mikaelian and Korzukhin 1997; Jenkins et al. 2004) and Australia (Eamus et al. 2000; Grieron et al. 2000; Keith et al. 2000). In West Africa, available functions (Adekunle 2007; Akindele and LeMay 2006; Henry et al. 2010; Sawadogo et al. 2010; Mbaekwe and Mackenzie 2008; Onyekwelu 2004; Guendehou et al. 2012) suggest that the topic has not been comprehensively addressed yet in the region. Henry et al. (2011) have compiled available allometric equations for sub-Saharan Africa and already underscored that there are very few equations for the region. This comprehensive compilation, as well as other recent studies (Djomo et al. 2010; Ebuy et al. 2011; Ryan et al. 2011; Vieilledent et al. 2012; Colgan et al. 2013; Fayolle et al. 2013; Mugasha et al. 2013; Mate et al. 2014; Ngomanda et al. 2014) did not include the majority of tropical tree species growing in West Africa. A recent study by Guendehou et al. (2012) also highlighted the need to expand the development of biomass models to further regions and tree species.

The use of generic allometric equations could be an alternative for carbon stocks quantification, especially in the tropics characterised by a diversity of tree species. Surprisingly, to our knowledge, trees used to develop these generalised equations (e.g. Brown 1997; Baker et al. (2004); Chave et al. 2005) did not include major tree species from Africa. Furthermore, several studies already found that the application of the generalised equations to area outside the data domain for which they were developed resulted in high uncertainties associated with tree biomass and volume estimation (Clark et al. 2001; Basuki et al. 2009; Alves et al. 2010; Henry et al. 2010; Fonseca et al. 2012; Guendehou et al. 2012; Lima et al. 2012; Fayolle et al. 2013; Ngomanda et al. 2014). This finding supports the development and application of local and species-specific biomass and volume models to study dynamics of carbon stocks.

Most of the approaches used to develop biomass models involved destructive sampling of trees (Gibbs et al. 2007; Devi and Yadava 2009; Peichl et al. 2012). This approach does not seem appropriate in the current context of using forests to mitigate climate change (Guendehou and Lehtonen 2014), as it releases an important amount of carbon to the atmosphere. Also, it does not protect threatened species in forest ecosystems. Furthermore, biomass models are to be consistent with allometric scaling laws (West et al. 1997; Enquist et al. 1998; West et al. 1999) which suggest that the size influences nearly all of the structural, functional and ecological characteristics of organisms and that the tree characteristics, including diameter and height, would be good predictors of tree volume and biomass. Some models including variables such as crown cover (JAFTA 2000) as input, previously developed in Benin, were difficult to apply as these variables were not always readily available or easy to measure.

In Benin, the very few studies on allometric equations (Fonton et al. 2002; 2009; Adjolohoun et al. 2013) were carried out in ecosystems other than semi-deciduous forest except for Guendehou

et al. (2012). The main objective of this study was to develop volume and biomass models for eighteen native dominant tree species in a natural tropical semi-deciduous forest called Lama in Benin, using non-destructive sampling approach. Further objectives were to develop generic volume and biomass models and determine basic wood densities of the tree species. The study intends to test the hypothesis that species-specific models are preferred to generic models.

2 Material and methods

2.1 Study site

The study site was the Lama forest reserve, a semi-deciduous forest (Nagel et al. 2004) located in southern Benin between 6°55' and 7°00' N and 2°04' and 2°12' E (Fig. 1). The forest consists of 16 250 ha, including 4 777 ha of natural forest called 'Noyau central', entirely protected.

The climate in southern Benin is tropical humid. The monthly average temperatures vary between 25 and 29 °C. The highest temperature 38 °C was observed between February and March and the lowest one (15 °C) was recorded in December. The site falls within a tropical moist zone according to the IPCC regional climate classification scheme (IPCC 2006). The mean annual precipitation in the experimental site is 1 200 mm with monthly precipitations exceeding 100 mm in all months except for January, February and December. Two wet and two dry seasons are observed throughout the year. The principal rainy season occurs between mid-March and mid-July and the shorter rainy season between mid-September and mid-November.

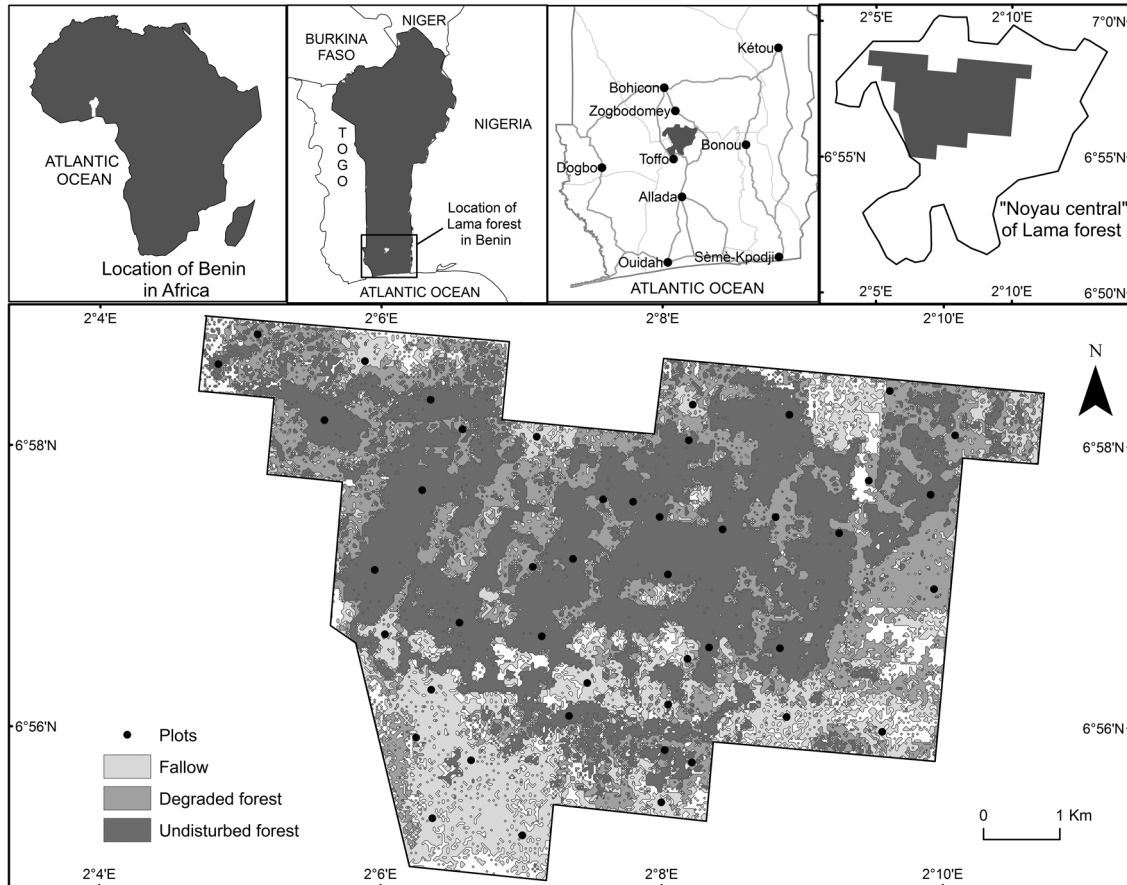


Fig. 1. Location of the study area.

The soil in the study area consists of hydromorphic clayey vertisol (40–60% of clay) characterized by a poor drainage and a pH range of 5–5.5 in the 0–30 cm horizon (Küppers et al. 1998). The pH increases up to 6.5–7 in deeper horizons due to the appearance of limestone at a depth of 150 cm.

Küppers et al. (1998) reported 67 families from a flora inventory carried out in the Lama forest. Mondjannagni (1969), Paradis and Hounnon (1977) and Akoègninou (1984) also described the species richness of the forest. Dominant tree species reported by previous studies (von Bothmer et al. 1986; Küppers et al. 1998; Nagel et al. 2004; Bonou et al. 2009) include indigenous species such as *Azelia Africana* Sm., *Anogeissus leiocarpa* (DC.) Guill. & Perr., *Ceiba pentandra* (L.) Gaertn., *Celtis brownii* Rendle, *Dialium guineense* Willd., *Diospyros mespiliformis* Hochst. Ex A.DC., *Holarrhena floribunda* (G. Don) Durand and Schinz, and *Mimusops andongensis* Hiern. The distribution of these species varies according to the vegetation cover (e.g. natural forest and fallow). The current study has extended the list of dominant tree species based on ground truth data as described below in section on species selection.

2.2 Sampling design and data collection

2.2.1 Plots selection

A systematic sampling procedure was first applied by overlaying a grid of points spaced from each other of 1 km on the study area map using QGIS 1.8 together with data from the most recent national forest inventory (NFI) carried out in 2007. The intersections of the lines on the grid were considered as the centre of the plots. Using this approach, 45 plots were identified within the boundary of the study area. Then, since the study area covers different vegetation types, the 45 plots were redistributed proportionately to the area of each vegetation type in order to avoid bias in the sampling. This resulted in 20, 10 and 15 plots of 50 × 50 m in natural dense forest, degraded forest and fallow respectively. Within each vegetation type, the allocated plots were redistributed randomly using the random point function of QGIS 1.8. The geographical coordinates of the centre of the plots were recorded using a GPS. The boundaries of the plots were defined using tape and a string. The geographical coordinates of the four corners of the plots were also recorded. These plots were used for the field measurements. The total area sampled was 11.25 ha and represented 0.23% of the “*Noyau central*” in line with the range reported by Chave et al. (2003; 2004) and GFOI (2013).

2.2.2 Species selection

Eighteen dominant tree species were selected based on the importance value index (IVI) computed following the approach of Curtis and Macintosh (1951). The IVI is used to determine the overall importance of each species in the community structure based on the relative frequency, relative density and relative dominance. These species included: *Terminalia superba* Engl. and Diels, *Lonchocarpus sericeus* (Poir.) Kunth, *Mimusops andongensis*, *Holoptelea grandis* (Hutch.) Mildbr., *Khaya senegalensis* (Desr.) A. Juss, *Albizia zygia* (DC.) J.F. Macbr., *Drypetes floribunda* (Müll. Arg.) Hutch., *Celtis brownie* Rendle, *Lecaniodiscus cupanioides* Planch. ex Benth, *Triplochiton scleroxylon* K.Schum., *Sterculia tragacantha* Lindl., *Diospyros abyssinica* (Hiern) F.White, *Malaecantha alnifolia* (Baker) Pierre, *Zanthoxylum zanthoxyloides* (Lam.) Zepern. & Timler, *Cynometra megalophylla* Harms, *Ficus sur* Forssk., *Cassipourea congoensis* R.Br. ex DC., *Holarrhena floribunda*. Other species with high IVI including *A. africana*, *A. leiocarpa*, *C. pentandra*, *D. guineense* and *D. mespiliformis* were already studied by Guendehou et al. (2012).

Table 1. Compilation of measurements of diameter and height for each sampled tree.

Species	Number of sample trees	Number of diameter measurement	Dbh range (cm)	Stem height range (m)	Total height range (m)
<i>Lonchocarpus sericeus</i>	23	145	3.2–38.5	2.28–20.48	2.28–24.15
<i>Drypetes floribunda</i>	28	162	2.4–31.5	2–9.96	2–12.09
<i>Albizia zygia</i>	28	168	2.5–65.5	3.03–26.73	3.03–28.08
<i>Lecaniodiscus cupanioides</i>	22	118	3.3–29.3	3.67–13.86	4.1–15.82
<i>Ficus sur</i>	27	163	4.4–67.7	3.25–21.85	3.25–26.64
<i>Zanthoxylum zanthoxyloides</i>	19	119	4.2–31.6	3.42–10.34	4.33–12.06
<i>Mimusops andongensis</i>	26	164	3.9–46.4	4–20.7	4–23
<i>Celtis brownii</i>	27	159	3.3–29.3	2.31–14.56	2.31–16.38
<i>Sterculia tragacantha</i>	31	194	2.9–34.6	2.02–16.2	2.02–18.8
<i>Cassipourea congoensis</i>	28	143	2.8–22.4	3.22–18.7	3.22–22.66
<i>Khaya senegalensis</i>	27	160	3.7–40.2	3.05–19.5	4.35–24
<i>Holarrhena floribunda</i>	32	187	2.9–41.3	2.1–20.25	2.1–23.25
<i>Cynometra megalophylla</i>	19	122	2.9–36	5.35–19.32	5.35–21.28
<i>Malacantha alnifolia</i>	27	169	2.9–31.9	3.6–13.72	3.6–17.92
<i>Terminalia superba</i>	37	240	8.2–44.4	6.32–26.04	9.44–33.93
<i>Triplochiton scleroxylon</i>	48	309	2.3–47.5	2.4–25.44	2.4–30.52
<i>Diospyros abyssinica</i>	18	109	1.8–32.7	3–19.26	3–22.14
<i>Holoptelea grandis</i>	34	207	3.5–32.9	4.62–22.54	5.4–25.3

2.2.3 Measurements of tree characteristics of individual trees

All living trees in each plot were measured for Dbh, using a non-destructive sampling and data collected were combined to calculate the IVI. Then, for the selected dominant species, detailed measurements on diameter along the stem and height were undertaken and wood samples were collected for basic wood densities measurement in laboratory. For studied species, each individual tree identified during the sampling was systematically measured for Dbh, and the data collected showed that at least one (1) tree was sampled in each existing Dbh class. The width of the diameter classes was 5 cm. Hollow trees and trees with broken top were not measured. Measurements were carried out between October 2013 and February 2014. In total 501 trees were measured. Table 1 gives the number of trees sampled, data on stem diameter, stem height and total height and the range of the measurements.

2.2.4 Measurement of sample trees

On each tree sampled, diameter measurement was performed first at the bottom (diameter on the ground), then at 1.30 meter above the ground and every meter along the stem from 1.3 m up to 6.3 m using a ladder. The absence of buttress on the sampled trees made it easier to measure diameter at the bottom. When the stem height was higher than 6.3 m, the diameter at the top of the stem was computed using a linear extrapolation based on two last diameters recorded along the tree stem (Guendehou et al. 2012).

Stem height is defined as height from the bottom to crown base. Using an optical Suunto clinometer PM-5, stem height and total height were estimated from observations recorded on the bottom, the crown base and the top of trees. Given the difficulty to record observations at the top of the tree in closed canopy, we assumed that total height measurement was uncertain and therefore was not used as variable in the model development.

Table 2. Mean wood density (g cm^{-3}) of the selected tree species in Lama forest reserve and comparison with other published data; wood density range includes all observations without modification. Wood density is given as oven-dry mass per fresh volume.

Species	This study				Others studies			
	Number of wood sample	Wood Density range	Mean (Standard deviation)	Coefficient of variation (%)	Brown 1997	Kindt et al. 2015	IPCC 2006	Zanne et al. 2009
<i>C. megalophylla</i>	34	0.81–1.3	0.98 (0.08)	7.96	-	-	-	-
<i>D. abyssinica</i>	31	0.45–1.90	0.86 (0.29)	33.73	-	0.83	-	-
<i>Z. zanthoxyloides</i>	33	0.65–1.43	0.84 (0.14)	16.62	-	-	-	-
<i>D. floribunda</i>	46	0.51–1.07	0.77 (0.12)	15.67	-	-	-	-
<i>M. andongensis</i>	48	0.42–0.98	0.77 (0.11)	14.51	-	-	-	-
<i>L. cupanioides</i>	36	0.42–1.03	0.77 (0.14)	17.86	-	-	-	-
<i>C. congoensis</i>	39	0.62–0.83	0.75 (0.05)	6.64	-	-	-	0.66
<i>L. sericeus</i>	42	0.48–0.95	0.75 (0.11)	14.15	-	0.70	-	0.75
<i>C. brownii</i>	42	0.52–0.98	0.73 (0.09)	13.04	-	0.72	-	-
<i>A. zygia</i>	52	0.44–0.73	0.65 (0.08)	13.31	0.46	0.51	-	0.49
<i>H. grandis</i>	58	0.42–0.76	0.63 (0.09)	14.08	0.59	0.61	0.59	0.59
<i>M. alnifolia</i>	46	0.48–0.73	0.61 (0.06)	9.24	0.45	-	0.45	-
<i>K. senegalensis</i>	52	0.41–0.74	0.59 (0.07)	11.83	0.60	0.66	-	0.63
<i>T. superba</i>	74	0.37–0.69	0.56 (0.06)	10.74	0.45	0.46	0.40–0.66	0.46
<i>H. floribunda</i>	57	0.43–0.61	0.54 (0.04)	8.32	-	0.47	-	0.47
<i>T. scleroxylon</i>	86	0.35–0.6	0.46 (0.07)	15.11	0.32	-	0.28–0.44	0.33
<i>F. sur</i>	50	0.32–0.52	0.45 (0.05)	11.97	-	-	-	-
<i>S. tragacantha</i>	47	0.16–0.44	0.32 (0.08)	25.80	-	-	-	-

For each sample trees, two wood samples were extracted at 1.3 m, at two points diametrically opposed, using an increment borer, except for trees in lower diameter classes ($\text{Dbh} < 5 \text{ cm}$) for which one sample was collected. Based on the diameter of the increment borer (5 mm) and the length of the fresh sample collected, the volume of the sample was estimated. In total, 873 wood samples were collected and oven-dried at 90°C to constant weight (during 48 hours) in laboratory. Dry mass of samples was measured using an electronic balance (Ohaus Pioneer Analytical Model scale) and the basic wood density was calculated as ratio of dry mass to fresh volume of the sample. Table 2 shows the basic wood densities of sampled tree species and the comparison with existing data.

2.2.7 Stem volume and biomass estimation

The stem volume of each tree was calculated as the sum of the volumes of sections of the stem, starting from the base up to the top of the stem (between 0 and 1.3 m; 1.3 m and 2.3 m and so on). The formula of a truncated cone (Netshiluvhi and Scholes 2001) was used to compute the volumes.

$$V = (\pi h / 12) \times (d_1^2 + d_2^2 + (d_1 \times d_2)), \quad (1)$$

where V is the volume of a section, $\pi = \text{pi}$, h = height of the stem section, d_1 and d_2 are the diameters of the truncated cone.

Stem biomass was calculated by multiplying the basic wood density with total stem volume.

2.3 Data analysis and model approaches

Volume and biomass models were developed and tested using the approach reported by Guendehou et al. (2012). Models were estimated and tested using the statistical computing software R (R Development Core Team 2012). A number of linear and non-linear volume and biomass functions, using only *Dbh* as variable or a combination of *Dbh* and stem height, were tested. A non-linear model form with multiplicative error was used as the basis of the model formulation. A logarithmic transformation was used to obtain homoscedastic variance, and to transform the equation to a linear form (Fig. 2).

The statistical parameters analysed for model comparison and selection included the standard error and significance of model parameters, the residual standard deviation of the model, the adjusted coefficient of determination, residuals vs. fitted, and the relative root mean square error (rRMSE). The analysis resulted in the selection of the following models:

$$\ln(X) = x_0 + x_1 \ln(Dbh) + x_2 \ln(H) + \frac{\sigma^2}{2} \quad (2)$$

$$\ln(X) = x_0 + x_1 \ln(Dbh) + \frac{\sigma^2}{2} \quad (3)$$

where X = stem biomass (kg) or volume (10^{-3} m^3), Dbh = the diameter at breast height at 1.3 m (cm) above ground, H = stem height (m), x_0 , x_1 and x_2 are model parameters, \ln is the natural logarithm, and σ is the residual standard deviation of the linearized models.

These linear models were derived from non-linear model form to address the heteroscedasticity (non-constant variance) of observations (Vieilledent et al. 2012; Guendehou et al. 2012; Mc Roberts et al. 2015). Biomass and volume are derived using the following equations:

$$X = e^{\left[x_0 + \frac{\sigma^2}{2} + x_1 \ln(Dbh) + x_2 \ln(H) \right]} \quad (4)$$

$$Y = e^{\left[x_0 + \frac{\sigma^2}{2} + x_1 \ln(Dbh) \right]} \quad (5)$$

where X = biomass (kg) or volume (10^{-3} m^3), Y = biomass (kg) or volume (10^{-3} m^3), Dbh = the *Dbh* at 1.3 m (cm), H = stem height (m), x_0 , x_1 and x_2 are model parameters, and σ is the residual standard deviation of the linearized models.

Species-specific models were developed by fitting equations to species-specific observations on *Dbh*, stem height, volume and biomass only. For the generic models, all observations collected in this study and those from Guendehou et al. (2012) were combined. This resulted in 617 trees used for the generic volume and biomass models. Chave et al. (2004) recommended more than 100 trees. Generic models in this study were compared to specific models and other published generic models by analysing the differences between predictions and observations.

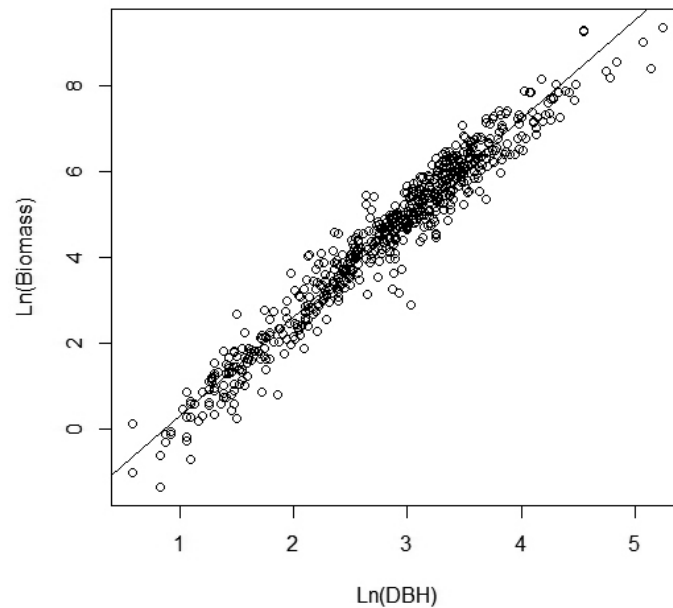


Fig. 2. Linear relationship with constant variance between \ln (biomass) and \ln (Dbh).

3 Results

3.1 Basic wood density

The highest value of basic wood density (1.90 g/cm^{-3}) was obtained for *D. abyssinica* whereas the lowest value (0.16 g/cm^{-3}) corresponded to *S. tragacantha* (Table 2). The highest mean value (0.98 g/cm^{-3}) was observed in *C. megalophylla* which is three times higher than the lowest mean density in *S. tragacantha*. *Drypetes floribunda*, *Mimusops andogensis* and *Lecaniodiscus cupanioides* had the same mean value (0.77 g/cm^{-3}) although the density values for *L. cupanioides* were distributed throughout a wider range. *Cassipourea congoensis* and *Lonchocarpus sericeus* also had the same mean value (0.75 g/cm^{-3}) but basic wood density data were distributed in a narrow range. The analysis of the observations indicated that only six out of eighteen tree species studied had basic wood density higher than 0.75 g/cm^{-3} (Table 2).

The coefficient of variation (CV) associated with species-specific basic wood density was less than 18% for all species except for *D. abyssinica* (34%) and *S. tragacantha* (26%) indicating low variation of wood density between individuals of the same species. However, lower values of densities were observed in younger trees in lower diameter classes. The variation of wood density between studied species was relatively high as indicated by the CV (30%).

3.2 Stem volume and biomass models

For both model formulation including models with Dbh only and models using a combination of Dbh and height as variables, high adjusted coefficient of determination was observed (Tables 3 and 4) suggesting that all species-specific and generic models accounted for a high proportion of variation of data. Inclusion of tree stem height in addition to stem diameter improved the performance of all models (lower rRMSE and residual standard deviation). Most of the model parameters were significant at $\Pr(>|t|) < 0.001$.

Table 3. Total stem volume models, parameter estimates, standard errors of parameter estimates, adjusted coefficients of determination, relative RMSEs and bias correction. V = Volume (10^{-3} m³); Dbh = the Dbh at 1.3 m (cm); H = tree stem height (m); x_0 , x_1 and x_2 are model parameters; RMSE = root mean square error; σ = model residual standard deviation. Stem is described as height to crown base. Figures in brackets are standard errors of the parameters.

Species	Model parameters			Adjusted R ² (%)	Relative RMSE (%)	$\sigma^2/2$	df
	x_0	x_1	x_2				
Model structure: $\ln(V) = x_0 + x_1 \ln(Dbh)$							
<i>H. grandis</i>	-1.58*** (0.11)	2.36*** (0.04)		99.03	15	0.01	32
<i>T. scleroxylon</i>	-1.96*** (0.10)	2.48*** (0.03)		99.13	17	0.02	46
<i>C. congoensis</i>	-1.90*** (0.11)	2.53*** (0.05)		98.74	17	0.02	26
<i>D. abyssinica</i>	-1.92*** (0.21)	2.51*** (0.08)		98.28	24	0.04	16
<i>Z. zanthoxyloides</i>	-1.71*** (0.18)	2.37*** (0.07)		98.6	18	0.02	17
<i>A. zygia</i>	-1.83*** (0.13)	2.49*** (0.04)		99.3	18	0.02	26
<i>K. senegalensis</i>	-1.76*** (0.15)	2.33*** (0.05)		98.71	18	0.02	25
<i>M. andongensis</i>	-1.71*** (0.15)	2.35*** (0.05)		98.86	18	0.02	24
<i>C. brownii</i>	-1.80*** (0.16)	2.37*** (0.06)		98.18	23	0.02	25
<i>S. tragacantha</i>	-1.47*** (0.18)	2.31*** (0.06)		97.66	24	0.03	29
<i>C. megalophylla</i>	-1.52*** (0.13)	2.31*** (0.05)		99.28	14	0.01	17
<i>L. sericeus</i>	-1.81*** (0.14)	2.42*** (0.05)		99.13	17	0.02	21
<i>H. floribunda</i>	-1.63*** (0.09)	2.36*** (0.03)		99.39	14	0.01	30
<i>D. floribunda</i>	-1.53*** (0.13)	2.23*** (0.05)		98.53	19	0.02	26
<i>L. cupanioides</i>	-1.38*** (0.21)	2.24*** (0.09)		96.88	23	0.03	20
<i>M. alnifolia</i>	-1.46*** (0.09)	2.23*** (0.04)		99.34	13	0.01	25
<i>T. superba</i>	-1.52*** (0.12)	2.35*** (0.04)		99.06	11	0.01	35
<i>F. sur</i>	-1.55*** (0.13)	2.30*** (0.04)		99.22	15	0.01	25
Generic	-1.48*** (0.04)	2.29*** (0.01)		97.51	32	0.05	615
Model structure: $\ln(V) = x_0 + x_1 \ln(Dbh) + x_2 \ln(H)$							
<i>H. grandis</i>	-2.10*** (0.11)	1.95*** (0.07)	0.66*** (0.10)	99.56	11	0.01	31
<i>T. scleroxylon</i>	-2.33*** (0.07)	2.04*** (0.05)	0.64*** (0.07)	99.7	10	0.01	45
<i>C. congoensis</i>	-2.36*** (0.17)	2.15*** (0.12)	0.60** (0.18)	99.09	14	0.01	25
<i>D. abyssinica</i>	-2.60*** (0.26)	1.96*** (0.18)	0.91** (0.27)	98.94	19	0.03	15
<i>Z. zanthoxyloides</i>	-2.63*** (0.21)	2.30*** (0.04)	0.53*** (0.10)	99.46	11	0.01	16
<i>A. zygia</i>	-2.10*** (0.10)	2.15*** (0.07)	0.52*** (0.09)	99.69	12	0.01	25
<i>K. senegalensis</i>	-1.90*** (0.10)	1.89*** (0.08)	0.62*** (0.10)	99.46	11	0.01	24
<i>M. andongensis</i>	-2.17*** (0.10)	1.99*** (0.05)	0.64*** (0.08)	99.66	10	0.01	23
<i>C. brownii</i>	-1.95*** (0.11)	2.04*** (0.08)	0.48*** (0.09)	99.12	14	0.01	24
<i>S. tragacantha</i>	-1.75*** (0.11)	1.84*** (0.07)	0.69*** (0.09)	99.21	14	0.01	28
<i>C. megalophylla</i>	-1.97*** (0.27)	2.17*** (0.09)	0.35 (0.19)	99.38	12	0.01	16
<i>L. sericeus</i>	-2.03*** (0.11)	2.07*** (0.08)	0.50*** (0.11)	99.56	12	0.01	20
<i>H. floribunda</i>	-1.95*** (0.07)	2.11*** (0.04)	0.44*** (0.06)	99.75	9	0.00	29
<i>D. floribunda</i>	-2.06*** (0.10)	2.03*** (0.04)	0.59*** (0.07)	99.58	10	0.01	25
<i>L. cupanioides</i>	-2.14*** (0.18)	1.92*** (0.07)	0.72*** (0.12)	98.9	13	0.01	19
<i>M. alnifolia</i>	-1.85*** (0.17)	2.08*** (0.07)	0.39* (0.15)	99.46	12	0.01	24
<i>T. superba</i>	-1.93*** (0.12)	2.14*** (0.05)	0.39*** (0.08)	99.43	9	0.00	34
<i>F. sur</i>	-1.85*** (0.13)	2.15*** (0.13)	0.32*** (0.09)	99.47	12	0.01	24
Generic	-2.10*** (0.04)	2.00*** (0.02)	0.63*** (0.53)	98.71	24	0.02	614

Table 4. Total stem biomass models, parameter estimates, standard errors of parameter estimates, adjusted coefficients of determination, relative RMSEs and bias correction. X = Biomass (kg); Dbh = the Dbh at 1.3 m (cm); H = tree stem height (m); x_0 , x_1 and x_2 are model parameters; RMSE = root mean square error; σ = model residual standard deviation. Stem is described as height to crown base. Figures in brackets are standard errors of the parameters.

Species	Model parameters			Adjusted	Relative	$\sigma^2/2$	df
	x_0	x_1	x_2	R ² (%)	RMSE (%)		
Model structure: $\ln(X) = x_0 + x_1 \ln(Dbh)$							
<i>H. grandis</i>	-2.44*** (0.10)	2.51*** (0.04)		99.23	16	0.01	32
<i>T. scleroxylon</i>	-2.64*** (0.14)	2.44*** (0.05)		98.24	24	0.03	46
<i>C. congoensis</i>	-2.25*** (0.12)	2.56*** (0.06)		98.6	17	0.02	26
<i>D. abyssinica</i>	-2.06*** (0.29)	2.49*** (0.11)		96.74	34	0.08	16
<i>Z. zanthoxyloides</i>	-2.36*** (0.21)	2.55*** (0.08)		98.27	22	0.02	17
<i>A. zygia</i>	-2.63*** (0.17)	2.60*** (0.05)		98.92	23	0.03	26
<i>K. senegalensis</i>	-2.50*** (0.16)	2.40*** (0.06)		98.56	22	0.02	25
<i>M. andongensis</i>	-1.98*** (0.20)	2.35*** (0.07)		97.99	24	0.03	24
<i>C. brownii</i>	-2.30*** (0.18)	2.44*** (0.07)		97.89	23	0.03	25
<i>S. tragacantha</i>	-2.93*** (0.22)	2.40*** (0.08)		96.48	32	0.05	29
<i>C. megalophylla</i>	-1.68*** (0.16)	2.37*** (0.06)		98.99	17	0.02	17
<i>L. sericeus</i>	-2.39*** (0.17)	2.53*** (0.06)		98.76	21	0.02	21
<i>H. floribunda</i>	-2.33*** (0.11)	2.39*** (0.04)		99	18	0.02	30
<i>D. floribunda</i>	-2.09*** (0.16)	2.35*** (0.06)		98.15	22	0.03	26
<i>L. cupanioides</i>	-2.21*** (0.25)	2.47*** (0.10)		96.44	27	0.04	20
<i>M. alnifolia</i>	-1.91*** (0.12)	2.21*** (0.04)		98.94	17	0.02	25
<i>T. superba</i>	-2.38*** (0.16)	2.45*** (0.05)		98.31	16	0.01	35
<i>F. sur</i>	-2.52*** (0.17)	2.35*** (0.05)		98.68	21	0.02	25
Generic	-1.98*** (0.06)	2.30*** (0.02)		95.07	45	0.10	615
Model structure: $\ln(X) = x_0 + x_1 \ln(Dbh) + x_2 \ln(H)$							
<i>H. grandis</i>	-2.72*** (0.14)	2.28*** (0.09)	0.35516* (0.14)	99.35	14	0.01	31
<i>T. scleroxylon</i>	-3.15*** (0.11)	1.86*** (0.08)	0.86*** (0.10)	99.28	16	0.01	45
<i>C. congoensis</i>	-2.71*** (0.17)	2.21*** (0.13)	0.58** (0.19)	99.06	15	0.01	25
<i>D. abyssinica</i>	-2.81*** (0.41)	1.88*** (0.28)	1.00* (0.43)	97.45	33	0.06	15
<i>Z. zanthoxyloides</i>	-3.38*** (0.28)	2.48*** (0.59)	0.58*** (0.14)	99.14	14	0.01	16
<i>A. zygia</i>	-2.99*** (0.12)	2.14*** (0.08)	0.70*** (0.11)	99.55	15	0.01	25
<i>K. senegalensis</i>	-2.64*** (0.12)	1.97*** (0.10)	0.60*** (0.13)	99.2	15	0.01	24
<i>M. andongensis</i>	-2.43*** (0.20)	2.00*** (0.11)	0.63*** (0.16)	98.72	18	0.02	23
<i>C. brownii</i>	-2.46*** (0.13)	2.08*** (0.09)	0.52*** (0.11)	98.88	16	0.01	24
<i>S. tragacantha</i>	-3.23*** (0.18)	1.90*** (0.12)	0.74*** (0.15)	98.03	21	0.03	28
<i>C. megalophylla</i>	-2.50*** (0.28)	2.11*** (0.09)	0.63** (0.19)	99.36	13	0.01	16
<i>L. sericeus</i>	-2.58*** (0.17)	2.23*** (0.13)	0.43* (0.17)	99.02	18	0.01	20
<i>H. floribunda</i>	-2.73*** (0.10)	2.08*** (0.06)	0.55*** (0.09)	99.55	12	0.01	29
<i>D. floribunda</i>	-2.63*** (0.15)	2.14*** (0.06)	0.60*** (0.11)	99.11	16	0.01	25
<i>L. cupanioides</i>	-2.99*** (0.25)	2.14*** (0.10)	0.74*** (0.17)	98.16	19	0.02	19
<i>M. alnifolia</i>	-2.35*** (0.22)	2.03*** (0.09)	0.44* (0.19)	99.09	15	0.01	24
<i>T. superba</i>	-2.83*** (0.19)	2.21*** (0.08)	0.43** (0.13)	98.71	14	0.01	34
<i>F. sur</i>	-2.94*** (0.17)	2.14*** (0.07)	0.45*** (0.12)	99.16	16	0.01	24
Generic	-2.63*** (0.07)	1.99*** (0.03)	0.67*** (0.04)	96.35	39	0.07	614

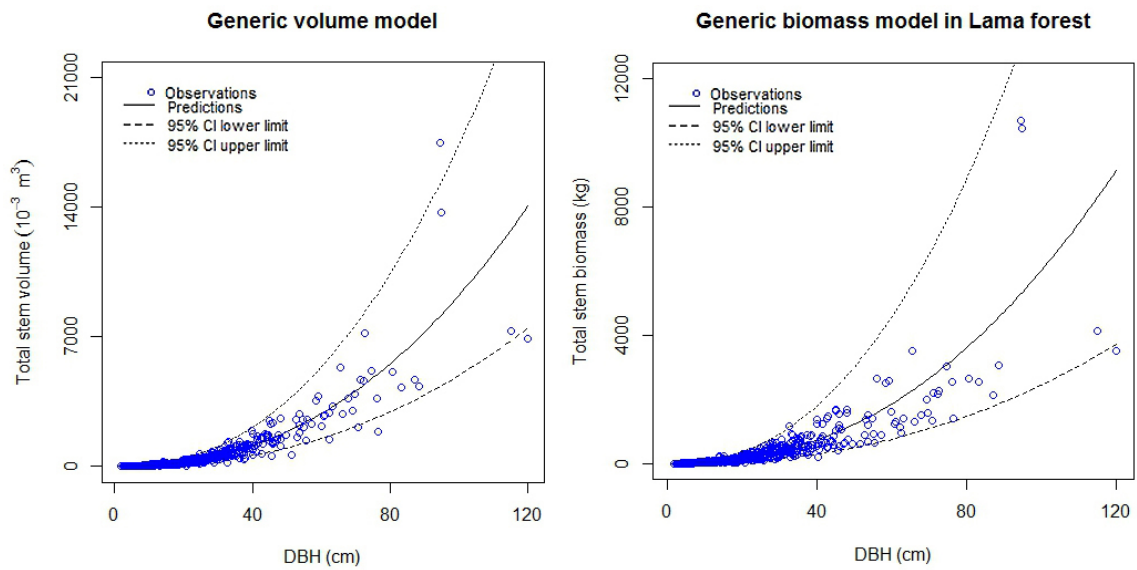


Fig. 3. Generic models for stem volume and stem biomass in Lama forest reserve.

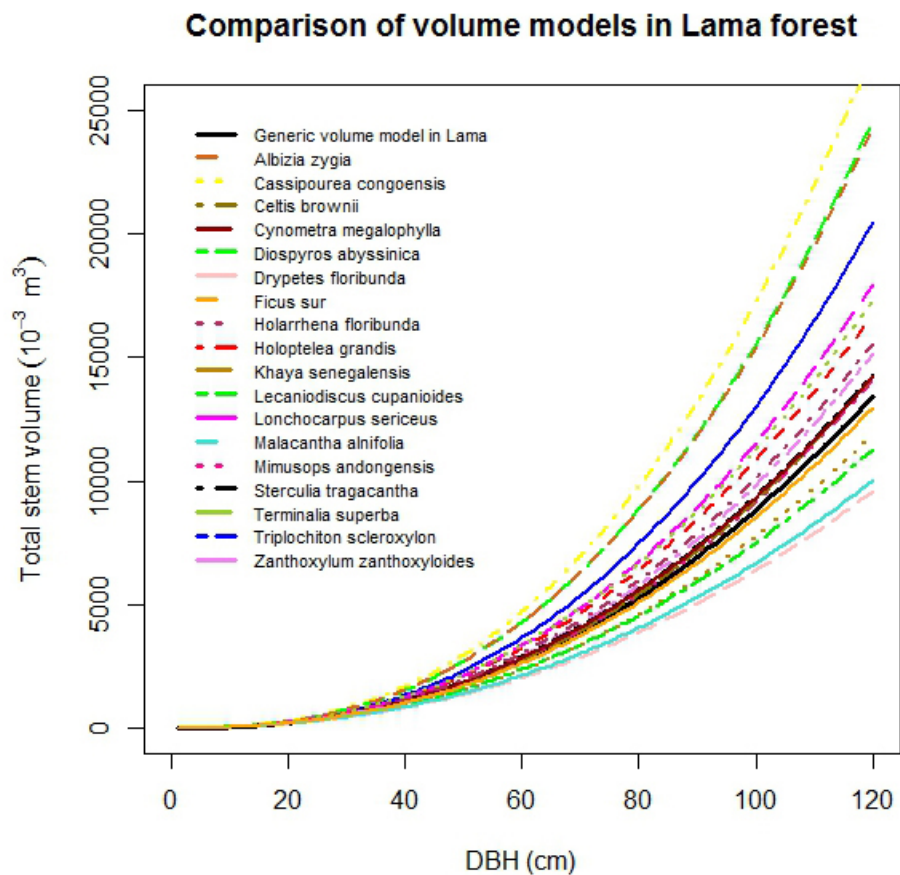


Fig. 4. Comparison of stem volume functions by selected tree species.

The fact that some observations were lacking in higher diameter classes ($Dbh > 50$ cm), due to the structure of the forest, (majority of the trees has $Dbh < 40$ cm) resulted in wide confidence intervals of the predictions (Fig. 3; Supplementary files 1 and 2, available at <http://dx.doi.org/10.14214/sf.1474>).

No large differences in stem volume among studied species were observed for Dbh less than 20 cm (Fig. 4). For Dbh more than 20 cm, the stem volume of *C. cassipourea* was the highest followed by *D. abyssinica* and *A. zygia*. The lowest stem volume was found for *D. floribunda*. The observed order of magnitude of stem volume was: *C. cassipourea* > *D. abyssinica* > *A. zygia* > *T. scleroxylon* > *L. sericeus* > *T. superba* > *H. grandis* > *H. floribunda* > *Z. zanthoxyloides* > *S. tragacantha* > *C. megalophylla* > *C. brownii* > *M. andongensis* > *F. sur* > *K. senegalensis* > *L. cupanioides* > *M. alnifolia* > *D. floribunda* (Fig. 4).

C. congoensis had the highest stem biomass. The combination of stem volume and wood density explained the observed order of magnitude of biomass in tree species: *C. congoensis* > *D. abyssinica* > *Z. zanthoxyloides* > *A. zygia* > *L. sericeus* > *C. megalophylla* > *L. cupanioides* > *H. grandis* > *C. brownii* > *T. superba* > *M. andongensis* > *D. floribunda* > *H. floribunda* > *T. scleroxylon* > *K. senegalensis* > *F. sur* > *M. alnifolia* > *S. tragacantha* (Fig. 5). For instance, although, *D. floribunda* had the lowest stem volume, its biomass was higher than in *S. tragacantha* which had the lowest stem biomass because of the lower wood density found in *S. tragacantha*. The generic biomass model provided good approximations for biomass estimation of *T. scleroxylon* and *K. senegalensis* only. Large differences were observed with the other species suggesting that the application of the generic model to these species will result in higher uncertainties.

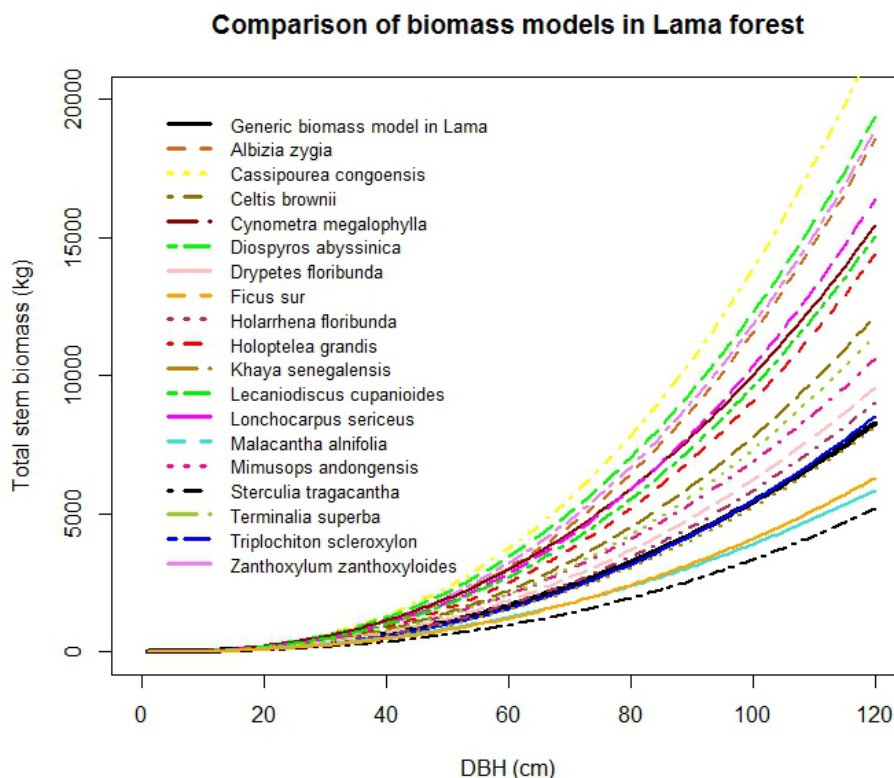


Fig. 5. Comparison of stem biomass functions by selected tree species.

4 Discussions

4.1 Basic wood density

The observed small differences in wood densities between individuals of same species are in agreement with the conservation features of wood density inside species reported by Chave et al. 2006; Swenson and Enquist 2007. Lower values of wood densities identified in general in younger trees in lower diameter classes were also reported by previous studies (Bao et al. 2001; Guendehou et al. 2012) who justified this finding by the fact that lignin content is lower in younger trees than mature ones. This finding also indicated that wood density in the studied species increases from pith to bark. As wood samples were not collected along the stem, this study could not conclude on the vertical variation of wood density. The non-destructive sampling applied was based on the assumption that wood density does not vary significantly from pith to bark along the stem. Given that tree species studied were in the same edaphoclimatic conditions, the large differences observed between wood densities of the eighteen tree species could not be directly attributable to environmental conditions, but more to differences between species (Nogueira et al. 2007; Gourlet-Fleury et al. 2011). These differences could be explained by the intrinsic characteristics of species (Chave et al. 2006; 2009; Zanne et al. 2010; Zeidler, 2012). Guendehou et al. (2012) already reported large differences in wood densities in the Lama forest. This finding is in agreement with Baker et al. (2004); Muller-Landau (2004) and Svob et al. (2014) who also identified that differences in wood densities in tropical forests are large.

Most of the tree species with low wood densities (less 0.75 g cm^{-3}) are fast-growing species except for *A. zygia*, *F. sur*, *M. alnifolia* and *S. tragacantha*. Similar results were found by Muller-Landau (2004) and Yeboah et al. (2014). In general, fast-growing trees are light demanding and produce low-density wood as found by Rueda and Williamson (1992), van Gelder et al. (2006) and Henry et al. (2010). However other studies have shown that some pioneer light demanding tree species can, unexpectedly, have high wood density (e.g *Lophira alata*, Fayolle et al. 2012) and this may be explained by wood composition and structure (Martínez-Cabrera et al. 2009; Poorter et al. 2010; Zanne et al. 2010).

For most tree species, mean values for wood densities reported in this study (Table 2) did not deviate significantly from data contained in published literature (Brown 1997; IPCC 2006; Zanne et al. 2009 and Kindt et al. 2015). For instance, mean value of 0.75 g cm^{-3} found in this study for *L. sericeus*, was also reported by Zanne et al. (2009) and was only 7% higher than the value reported by Kindt et al. (2015). Although, above consulted literature did not provide comprehensive information on the methodological approach used to determine densities, it seems that the reported densities were determined using volumetric approach. The fact that the results of this study were close to the published data justifies that the non-destructive sampling approach used to determine wood density was reliable. This comparison highlighted the double advantage of the non-destructive sampling used to determine wood density, including providing reliable estimates and protecting trees and forest ecosystems. Few studies reported wood densities based on non-destructive sampling (Chave et al. 2006; Sungpalee et al. 2009; Gryc et al. 2010; Carson et al. 2014) but none of them reported on the tree species in this study.

4.2 Stem volume and stem biomass models

Given that the studied species were in the same edaphoclimatic conditions, which means uniform forest, the model specification was based on the assumption that the between-plot and the within-plot variations of tree properties were not significant as confirmed by the analysis of the basic wood density.

Models with two variables (Dbh and height) performed well, for all species, compared to models that use Dbh as predictor, based on statistical analyses (higher R^2 , lower rRMSE and residual standard deviation). This finding was reported in other studies (Guendehou et al. 2012; Fayolle et al. 2013; Ngomanda et al. 2014) which indicated that the inclusion of height in addition to stem diameter improved the performance of models. Models with one variable had also good predictive ability and can be used in the absence of stem height. High adjusted coefficient of determination ($\text{Adj } R^2 > 95\%$) found for both types of models (one and two variables) highlighted that the models accounted for a high proportion of variability in predictions, indicating that most of the variance was explained by the models. This was partly due to the stratified sampling approach used that collected observations on tree size in each existing diameter class (Marková and Pokorný 2011; Cienciala et al. 2013; Guendehou et al. 2012; Sileshi 2014).

The generic biomass and volume models developed in this study performed well, for all species in lower diameter classes only ($\text{Dbh} < 20 \text{ cm}$), when compared to species-specific models (Fig. 4 and Fig. 5). For large trees, the generic biomass model provided good predictions for two species (*T. scleroxylon*, *K. senegalensis*) only out of eighteen. The model either overestimated (up to 60%) or underestimated (up to 63%) biomass for the other species (Fig. 5). Similar observations were made for the generic volume model which provided good predictions for *F. sur* and *Mimusops andongensis* only (Fig. 4). Given the importance of large trees in carbon budget in forest ecosystems, carbon stocks in large trees should be estimated with accuracy. But, the generic model was not able to predict biomass for the majority of studied species even in the same forest ecosystem (same edaphoclimatic conditions). This implies that specific models are preferred for an accurate volume and biomass estimation.

Brown (1997) developed generic model using observations collected in Africa, America and Asia to estimate total aboveground biomass. *Ceiba pentandra*, *Holoptelea grandis*, *Khaya senegalensis*, *Malacantha alnifolia*, *Terminalia superba* from our study were also included in the database used by Brown (1997) who provided up to 316% higher biomass estimates for all species except for *C. congoensis* for which biomass estimate was 2% lower (Supplementary file 3, available at <http://dx.doi.org/10.14214/sf.1474>). Guendehou et al. (2012) reported similar finding when applying Brown (1997) to *Azzeria africana*, *Anogeissus leiocarpa* and *Ceiba pentandra* in Lama forest, but they also found that Brown (1997) provided lower biomass estimates for *Diospyros mespiliformis* and *Dialium guineense*. This shed doubt on the applicability of generic model to estimate biomass.

The generic model developed by Djomo et al. (2010a) based on data collected from Brown 1997; Araújo et al. 1999; Nelson et al. 1999; Ketterings et al. 2001 and Djomo et al. (2010b), when applied to the observations of this study gave 2–62% lower biomass estimates for fourteen tree species studied and 3–62% higher biomass estimates for *K. senegalensis*, *S. tragacantha*, *M. alnifolia* and *F. sur* (Supplementary file 3). Even if the database used by Djomo et al. (2010a) was larger than that of Brown (1997), their generic model was not able to predict with accuracy biomass estimates for the tree species in our study. Similar conclusions can be formulated when comparing the generic model from Basuki et al. (2009) in Indonesia to our data. Basuki et al. (2009) seems to be applicable to *C. brownii*, *T. superba* and *M. andongensis* only. Predictions of biomass with this model, resulted in higher estimates up to 113% (*S. tragacantha*) and lower biomass estimates up to 50% (*C. congoensis*) while the model was supposed to give higher estimates because developed to generate total above ground biomass. The generic model from Ngomanda et al. (2014) in Gabon also provided 1–31% lower biomass estimates for *C. congoensis*, *D. abyssinica*, *Z. zanthoxyloides*, *A. zygia*, *C. megalophylla* and *L. sericeus* while it should be the opposite.

From the above discussions, one could confirm the hypothesis that it is difficult to identify which of the several existing biomass and volume equations would be suitable to apply in given edaphoclimatic conditions. Several factors may explain this inability of generic models to provide accurate estimates. Even if this study has not addressed the issue thoroughly, it suggests that among these factors, bio-physiological properties of species and edaphoclimatic conditions would have larger effects, suggesting that considering all species together to develop generic models and applying models to regions for which they were not developed would not be consistent with ecological laws.

Given that wood densities data generated in this study were in line with published data in one hand and due to the fact that, technically, there are no big differences between measuring stem diameter and height (and thus stem volume) on standing trees and deriving these variables on logged trees, in another hand, justified the robustness of volume and biomass models developed in this study. Vann et al. (1998) already found that allometric equations developed from non-destructive sampling appears to be as robust as equations derived using actual weights.

To account for the aboveground biomass, the stem biomass estimated using the models developed in this study should be multiplied by the available biomass expansion factor (IPCC, 2006).

This study is among the first initiatives, in West Africa, using ground truth data collected from non-destructive approach on large tree sample to develop volume and biomass models. The development of the models is consistent with the higher tier method of the Intergovernmental Panel on Climate Change (IPCC 2003; 2006) to estimate carbon stocks and change in carbon stocks in forestry. The study contributes to expand the database on biomass and volume models and basic wood densities in Africa (Henry et al. 2011). It makes available data and tools to countries in the region to assist them to report their contribution to the global effort to mitigate climate change through the implementation of activities such as REDD (reducing emissions from deforestation and forest degradation, conservation of forest carbon stocks, sustainable management of forests, and enhancement of forest carbon stocks in developing countries). The study would also assist countries to meet their reporting requirements under the United Nations Framework Convention on Climate Change (UNFCCC) through the submission of greenhouse gas (GHG) inventories in biennial update reports and national communications (UNFCCC 2010).

5 Conclusions

The logarithmic model using Dbh and height as independent variables was the most suitable for volume and biomass estimation for the eighteen tree species studied. In circumstances where stem height is not available, the model with Dbh only can be used as it gives also good predictions. Based on the comparison between species-specific models and the generic model from the same site and with generic models in published literature, this study concludes that species-specific volume and biomass models are more appropriate for estimating tree volume and biomass. This work is a significant contribution to database on wood densities and allometric equations in tropical countries. It demonstrated that the non-destructive sampling approach needs to be promoted as the wood densities it generated do not deviate significantly from those of destructive sampling and that the models developed based on data collected from this sampling are robust. The data and models generated in this study are suitable for reporting carbon stocks and its changes under REDD, Clean Development Mechanism of the UNFCCC.

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Total of 92 references.

Supplementary files

Available at <http://dx.doi.org/10.14214/sf.1474>:

- Supplementary file 1: Species specific models for stem Biomass in Lama forest reserve [Figures]
 Supplementary file 2: Species specific models for stem volume in Lama forest reserve [Figures]
 Supplementary file 3: Comparison of stem biomass function by generic models [Figures]